

Species diversity, structure and dynamics of two populations of an endangered species, *Magnolia dealbata* (Magnoliaceae)

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Abstract: Little is known about the ecology and demography of the genus *Magnolia*. *Magnolia dealbata* Zucc. is an endangered species endemic to Mexico. Two contrasting populations of *M. dealbata* (one from the grasslands and other from a secondary cloud forest) were studied. We asked the following questions: (a) Are size structure (diameter at breast height, DBH) and infructescence production significantly different between the two populations? (b) What are the populations' growth rates (λ) based on an initial 1987 study? (c) Are the associated species diversity indices of these *M. dealbata* populations significantly different? The results show no significant differences between the population size structure ($p=.094$); the growth rates of the populations were 0.992 in grassland and 1.053 in secondary cloud forest. The number of infructescences produced in year 2001 and DBH relationship were significantly linear ($p<.001$) in both populations, and there was no significant difference ($p>.01$) between their slopes. The diversity indices were not significantly different ($p>.05$), and only 54% of the species were common to both sites. Our study suggests that both populations are relatively stable and that the management history could impact more on the species composition than on the diversity indices. *Rev. Biol. Trop.* 54 (3): 997-1002. Epub 2006 Sept. 29.

Key words: cloud forest, diversity, endangered species, infructescences, *Magnolia dealbata*, Mexico, population.

Few population studies have been done on the genus *Magnolia* (eg Weaver 1987, Doyle 1989, Gutiérrez 1993). Seven species and two subspecies of *Magnolia* are native to Mexico (Vázquez 1994): *M. iltisiana* and *M. schiedeana* are classified as threatened species, and *M. dealbata* and *M. sharpii* are classified as endangered according to IUCN and Federal Mexican laws. *Magnolia dealbata* is endemic to Mexico and it is actually restricted to small local populations in the states of Oaxaca, Hidalgo, San Luis Potosí and Veracruz. Coyopolan, Veracruz has the highest number of individuals (Pattison 1985, Gutiérrez 1993). A recent study reported a new locality for *M. dealbata* in the state of Querétaro, but it did

not describe the condition of the population (Cartujano *et al.* 2002).

M. dealbata wood is utilized as firewood and fence material. Its giant white flowers are used ornamentally and for traditional medicine (Gutierrez and Vovides 1997). Indigenous Mexicans know *M. dealbata* as *Eloxochitl* (from nahuatl *elotl* ear and *xochitl* flower) (Pattison 1985). *M. dealbata* usually grows in cloud forest or "bosque mesófilo de montaña" (Rzedowski 1978).

Gutierrez and Vovides (1997) described the annual survivorship of two *M. dealbata* populations from Coyopolan, Veracruz, based on data from 1986 to 1987 (L. Gutiérrez, pers. com.). There are no other studies published on

the present population status of this rare species. Similarly, there are no descriptions or listings of the species associated with *M. dealbata*.

Information on population structure and size is useful, even from short-term studies, to allow us to make inferences about the past and future of a particular population (Martínez-Ramos and Álvarez-Buylla 1995, Sánchez-Velásquez *et al.* 2002). Our study is a preliminary investigation of a long-term study in which the general objective is to document the perturbation impact on the demography of *M. dealbata* in Coyopolan, Veracruz, México. Two ecologically contrasting populations of *M. dealbata* were studied (grassland and secondary cloud forest) with the following questions to test: (a) Are the size structure (DBH) and infrutescences production significantly different in both populations? (b) What are the populations' growth rates (λ) based on the initial study of Gutiérrez from 1986-1987? (c) Are the associated species diversity indices of the *M. dealbata* populations significantly different?

MATERIALS AND METHODS

M. dealbata, a member of Section Rytidospermum, is a deciduous tree that grows to 25 m in height, has white large flowers measuring 20 cm and leaves 80 cm long. Each adult tree is able to produce several new stump sprouts after they have been cut, but only one or two principal stems will survive. These attributes produce no clear relationship between age and size, as we found an old tree with a small stem as well as an old tree with a large stem. Altitudinal distribution of *M. dealbata* is from 1200 to 1500 m, Pattison (1985), Johnson (1989), Vázquez (1994), Qiu *et al.* (1995), Vovides and Iglesias (1996), Gutiérrez and Vovides (1997), Azuma *et al.* (2001) and Kim *et al.* 2001, for more details of its ecology, taxonomy and phylogeny.

We studied two populations of *M. dealbata* in Coyopolan, Veracruz, México (19°21' N, 97°04' W), one located in secondary cloud forest and the other in grassland. Both were

previously studied during 1986-1987 by Gutiérrez (1993), but we did not find the exact location of the originals plots. The climate in this area is subtropical, with rainfall of 1 500-2 000 mm annually distributed throughout the year. The average annual temperature is 19.5 °C (Gutiérrez and Vovides 1997). Eight random plots (10 x 10 m) were used for each population (a total of 16), accounting for all individuals of *M. dealbata*. Other species included were only those ≥ 1.3 m in height; species, DBH and height were registered. Our plot size was larger than the size of the largest cluster (6 x 5 m) of *M. dealbata* found (Gutiérrez 1993). Number of infrutescences produced by individuals of *M. dealbata* was counted. Each individual *M. dealbata* seedling was mapped and labeled for a subsequent demography study.

The size structure between both populations was analyzed by Spearman correlation coefficient (Zar 1999). Dependence between size classes of *M. dealbata* vs. sites was determined by Fisher test exact.

The χ^2 test was used to find independence (total number of individuals of *M. dealbata*) between sites and year (years 1987 and 2002). Previously, the registered data was recorded as the number of individuals per hectare. The annual population growth rate (λ) was calculated in both populations by using the equation:

$$N_t = N_0 \lambda^t,$$

where N_t is population size in time t (until year 2001), N_0 is initial population size (year 1987), λ is population growth rate and t is time (14 years). Hence λ value can be estimated by

$$\lambda = \sqrt[t]{\frac{N_t}{N_0}}$$

The infrutescences number vs. DBH size was analyzed by linear simple regression in both *M. dealbata* populations (DBH as independent variable).

Shannon-Wiener diversity indices were calculated for both sites using the equation:

$$H' = -\sum_{i=1}^s (p_i)(\log_2 p_i),$$

where s is species number; p_i is proportion of species i -th. The diversity indices were compared with Student's T test (Zar 1999, Sánchez-Velásquez and Pineda-López 2000). We used individuals > 1.3 m height. The structure of plant species community associated with *M. dealbata* was analyzed through χ^2 test.

RESULTS

The size structure between both populations was not significantly correlated (Fig. 1) (Spearman, $p=.094$), nor was size structure *vs.* sites significantly dependent (Fisher, $p=.276$).

In the 1986-1987 study, the number of individuals of *M. dealbata* per hectare was 338 and 554 in grassland and secondary cloud forest, respectively (Gutiérrez 1993), while in this study (2001) was 300 in grassland and 1 137 in secondary cloud forest. The total number of individuals of *M. dealbata* between sites and years was not independent (χ^2 , $p<.001$). The population growth rate in grassland was 0.992, while in secondary cloud forest was 1.053.

Linear regression analysis between relationship of the number of infructescences and DBH was significant in both populations; grassland ($Nc = -7.28+1.556$ DBH; r^2 , $p<.001$) and secondary cloud forest ($Nc = -5.11+1.549$ DBH; r^2 , $p<.001$). However, the slopes were not significantly different between populations (Student's T, $p>.1$).

The diversity index was not significantly different between sites (Student's T, $p>.05$; secondary cloud forest $H' = 2.62$ and grassland $H' = 2.84$). Nevertheless, only 54 % of the species were common to both sites (year 2001). Five species from grassland were not found in secondary cloud forest, while two species from secondary cloud forest were not established in grassland (Fig. 2).

DISCUSSION

The size structure in both *M. dealbata* populations was an inverse *J* type. This is

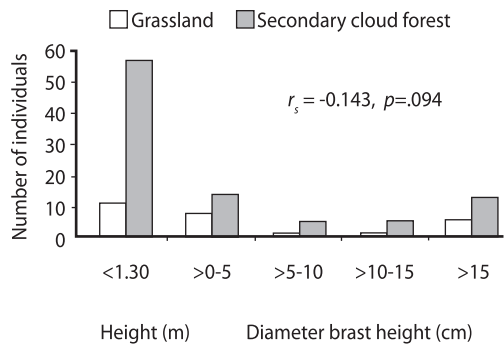


Fig. 1. Size structure of two *Magnolia dealbata* populations in Coyopolan, Ixhuacán de los Reyes, Veracruz, Mexico.

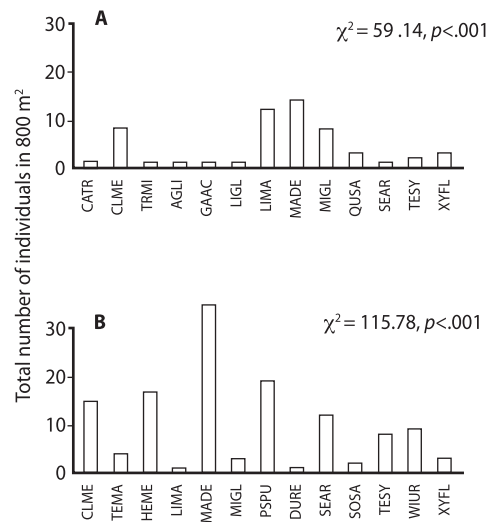


Fig. 2. Floristic composition associated with *Magnolia dealbata* (MADE), individuals > 1.3 m tall. a) Grassland, b) Secondary cloud forest. *Carpinus tropicalis* (CATR), *Clethra mexicana* (CLME), *Trema micrantha* (TRMI), *Ageratina ligustrina* (AGLI), *Gaultheria acuminata* (GAAC), *Litsea glaucescens* (LIGL), *Liquidambar macrophylla* (LIMA), *Miconia glaberrima* (MIGL), *Quercus sartorii* (QUSA), *Senecio arborecens* (SEAR), *Ternstroemia sylvatica* (TESY), *Xylosma flexuosum* (XIFL), *Tetrachyron manicatum* (TEMA), *Hedyosmum mexicanum* (HEME), *Psychofria pubescens* (PSPU), *Solanum salvifolium* (SOSA), *Wigandia urens* (WIUR), *Duranta repens* (DURE).

typical for many natural populations of tree species (Vanclay 1994) and shade tolerant species. This type of curve, in which age and size are not correlated, suggests that there are more environmentally poor conditions than rich for the growth of individuals (Martínez-Ramos and

Álvarez-Buylla 1995). However, we infer in our study that this is not the case: from almost all the individuals that had been cut recently at the base emerged more small individuals than large. As mentioned earlier, *M. dealbata* can sprout several stems after it has been cut, but only one or two principal sprouts survive (pers. obs.). We could identify sprouts because some individuals had evidence of cuttings. Similarly, the size structure of the population (J inverse and non relationship between age-size) of *Magnolia splendens* was observed in the Luquillo mountains of Puerto Rico (Weaver 1987) and also in *M. macrophylla* in the Mill Creek Forest, USA (Doyle 1989).

In the secondary cloud forest the population growth rate (λ) of *M. dealbata* was estimated at a 5.4% annual population increase, ie $\lambda = 1.054$. This suggests a good history of *in situ* conservation management and also $\lambda > 1$ is typical of a shade tolerant species (Vanclay 1995). Nevertheless, the grassland population of *M. dealbata* (an open area and frequently used for cow grazing) showed a slight annual population decrease, a 1% annual loss. However, more details and time are necessary to understand the perturbation impact on population dynamics of *M. dealbata*. Weaver (1987) found inversed results: he registered an 18% population increase of *M. splendens* in clear cut sites, while in sites without clear cutting, he found a *M. splendens* population decrease of 35% in 35 years. There have been other studies in Mexico of two evergreen species; briefly, *M. sharpii* is a local endangered species in Los Altos de Chiapas (González *et al.* 1997) and *M. iltisiana* is a threatened species with low regeneration and few individuals in the Western Sierra de Manantlán of Mexico (Vázquez *et al.* 1995, Saldaña-Acosta *et al.* 2001).

The significant linear regression of DBH vs the infrutescences number within each of the populations and the lack of significant difference between the slopes implies that *M. dealbata* individuals produce more infrutescences as size increases. This characteristic is common in nature because the tallest

individual captures more resources than the shorter individuals, and in turn produces more infrutescence. Future demography studies or conservation agencies can use these models (relationship DBH-Infrutescences number) to forecast future production of seeds.

We found more species simultaneously with *M. dealbata* (19) than the six cited in Gutiérrez and Vovides (1997), but we did not find *Tilia sp.* as they reported. Perhaps, we found more species because we used smaller size DBH than they did.

Diversity index comparisons of the different studies have some limitations (Sánchez-Velásquez *et al.* 2002b) because: 1) different sample size, plot size and diversity index are used, and 2) the minimum size of the individuals used are different. However, with these limitations we can observe that our index values are in the intervals of those found in other studies of Mexican cloud forests, but our values are higher than their average values: Sierra de Manantlán Biosphere Reserve (H' from 3.2 to 0.59, with 5 cm of DBH minimum) (Santiago and Jardel 1993, Hernández *et al.* 2000), El Cielo Biosphere Reserve (H' from 3.6 to 0.39, with 5 cm of DBH minimum) (Puig 1987), Sierra de Cucharas (H' from 3.89 to 2.81, with 15 cm of DBH minimum) (Puig *et al.* 1983) and El Triunfo ($H' = 3.8024$, with 5 cm of DBH minimum) (Williams 1991).

Our study suggests that a change in management history can impact species composition more than diversity indices. Only 54% of the species were common to both plant communities associated with *M. dealbata*.

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RESUMEN

La ecología y demografía del género *Magnolia* han sido poco estudiadas. *Magnolia dealbata* Zucc. es una especie endémica de México y es considerada en peligro de extinción. Se estudiaron dos poblaciones contrastantes de *M. dealbata* (una dentro de un pastizal y la otra dentro de vegetación secundaria del bosque mesófilo de montaña –bosque de niebla-). Buscamos respuestas a las siguientes preguntas: (a) ¿La estructura de tamaños y la producción de frutos de ambas poblaciones son significativamente diferentes? (b) ¿Cuáles son las tasas de crecimiento poblacional (λ), tomando como población inicial un estudio de 1987? (c) ¿Son significativamente diferentes los valores de diversidad de especies de ambas poblaciones? No hallamos diferencias significativas entre la estructura de tamaños de ambas poblaciones ($p=0.94$). Las tasas de crecimiento poblacional de las poblaciones en pastizal y en el bosque secundario de neblina fueron 0.992 y 1.053, respectivamente. El número de infrutescencias por individuo producidas en 2001 y su relación con el diámetro a la altura del pecho fueron significativamente lineales, y no hubo diferencias significativas entre sus pendientes ($p>.1$). Los índices de diversidad no fueron significativamente diferentes ($p>.05$) y sólo el 54 % de las especies fueron comunes entre sitios. Las poblaciones parecen relativamente estables y el historial de manejo en ambas comunidades podría afectar más a la composición de especies que a los índices de diversidad.

Palabras clave: bosque nuboso, diversidad, especie en peligro de extinción, infrutescencia, *Magnolia dealbata*, México, población.

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