

Is forest cover conserved and restored by protected areas?: The case of two wild protected areas in the Central Pacific of Costa Rica

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Abstract: Changes in land use are mainly a consequence of anthropogenic actions. The current agricultural and urban transformations in Costa Rica have raised questions about the effectiveness of conservation and restoration within protected areas. Herein we analyzed the patterns of land use change between three periods: 1997, 2005 and 2010 in terms of magnitude, direction, and pace through categorical maps generated by the photo-interpretation for La Cangreja National Park (LCNP), Rancho Mastatal Wildlife Refuge (RMWR), and their surrounding areas (SA), this last compound of one kilometer radius outside the protected areas' boundaries. The matrix which describes the landscape within the protected areas is natural coverage, composed mainly by forest cover and thickets. We found that the most abundant natural cover for both protected areas was forest cover for all years tested. The stability and large areas of forest cover in LCNP and RMWR for 2005 and 2010, reflected that policies, management actions and vigilance, have a positive impact on the conservation and restoration of natural habitats in these Costa Rican Central Pacific areas. However, the high landscape complexity of the SA in 1997, 2005 and 2010 was an evidence of the anthropogenic pressure on these protected areas, and suggested the ineffectiveness of local governments to monitor and abate land use changes, that could hinder the management, conservation and restoration of species in the protected areas. *Rev. Biol. Trop.* 63 (3): 579-590. Epub 2015 September 01.

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Changes in land use in terms of magnitude, direction, and pace, are mainly consequences of anthropogenic actions (Aguayo, Pauchard, Azócar, & Parra, 2009; Bender, Boehmer, Jens, & Schumacher, 2005; Wiens, Stenseth, Van Horne, & Anker, 1993). In recent decades, agricultural and urban transformation have generated a landscape with heterogeneous characteristics (Rosero-Bixby & Palloni, 1998; Turner, 2005), which describes a mosaic of natural and human-driven spaces that define spatial and temporal patterns on lands (McLennan & Garvin, 2012; Peña-Cortés et al., 2006). From a conservation perspective, these heterogeneous landscapes result in loss, fragmentation and modification of forest habitats, which

leads these areas to have a decline in flora and fauna communities (Fahrig, 2003; Haila, 2002; Hanski, 1998; Sanchez-Azofeifa, Harriss, & Skole, 2001).

To prevent changes in land use in forested areas and conserve remnant tree species, various official organizations have created different categories of protected areas that restrict land use changes and deforestation (Myers, 1993; Sanchez-Azofeifa, Rivard, Calvo, & Moorthy, 2002). Since 1995, Costa Rica has a protected areas system that allows *in situ* conservation through different management categories that dampens, to some extent, land use changes in terrestrial ecosystems (SINAC, 2010). While it is true that the country has



generated many protected areas in the last two decades, most of these are currently immersed in a highly dynamic anthropogenic landscape matrix that hinders conservation objectives and management plans (Broadbent et al., 2012; Sánchez-Azofeifa, Daily, Pfaff, & Busch, 2003; Sanchez-Azofeifa et al., 2002). The protected areas of the Central Pacific in Costa Rica, specifically La Cangreja National Park (LCNP) and Rancho Mastatal Wildlife Refuge (RMWR) are a good example. Before their foundation in 2002 and 2004 respectively, both areas had suffered extensive deforestation, so much so that in 1980 the government declared the region an emergency zone to stop the overexploitation of resources (Bonilla, 1983). Despite multiple efforts, the region is still suffering from wildlife hunting, burning of land, urban growth and deforestation around them (Jiménez, 2011). Therefore, to visualize the dynamics of land use changes in and around these protected areas, to better focus future management efforts, it is important to determine which sites have been significantly affected and the possible reasons.

The aim of this paper was to analyze patterns of land use changes in LCNP, RMWR, and surrounding areas, during the period from 1997 to 2010 to be used as a technical-scientific tool in the management of protected areas. We hypothesized that protected areas are entities of conservation and restoration with increased coverage, connectivity and stability of forest cover compared to the surrounding areas. Specifically in this study, we investigated the magnitude and rate of land use change, the diversity and equity of the landscape elements, and finally characterized those regions that have experienced the greatest land use change and its origin and expansion direction.

MATERIALS AND METHODS

Study site: This study was conducted in two protected areas in the Central Pacific of Costa Rica: La Cangreja National Park (LCNP) and Rancho Mastatal Wildlife Refuge (RMWR) (Fig. 1). This region contains one of

the last remnants of montane rainforest in the Central Pacific (Costanzo, 2006; Holdridge, 1967). Due to high annual rainfall (4 000 mm), high humidity, nutrient-poor soils and varied topography (Jiménez, 1999), the area contains about 44 endemic tree species (Acosta-Vargas, 1998), which makes this area one of the highest of biodiversity and floristic endemism in the country. To delimit the study areas, we worked with three previously established areas: 1) LCNP limits established in 2005 (Poder Ejecutivo, 2005; 2 508.19 ha), 2) the RMWR limits established in 2004 (Poder Ejecutivo, 2004; 79.16 ha), and 3) the surrounding area (SA) (2 720 ha). The SA was obtained using a 1 km radius from the edge of the protected areas. The entire landscape in this study is within the Paso de Lapas biological corridor (Bustamante, 2006).

Cartographic process: We used 1:40 000 scale aerial photographs for 1997 from Costa Rican TERRA project, 1:20 000 scale aerial photographs for 2005 from Costa Rican CARTA project, and 1:40 000 scale satellite photographs from ArcGIS online to create thematic maps in each delimited area. Aerial and satellite photographs were georeferenced to CRTM05 coordinate system of Costa Rica, Datum WGS84. Georeferenced images were performed by ArcGIS 10 software through a georeferencing module used as reference digital map for the area and georeferenced and orthorectified photomosaic from national cadastre system of 2005. For each photograph, 40 control points were used to improve the accuracy of the georeferencing technique and a 3rd order polynomial transformation was used for rectification. From georeferenced image sets, photomosaics were made and then used as basis for the digitization of thematic layers.

The creation of the thematic layers was made from photo-interpretation of the photomosaic per year. From these, each element was considered unique in the landscape with a clearly identifiable color and characteristic structure. Photo-interpretation was made firstly by identification of which types of coverage

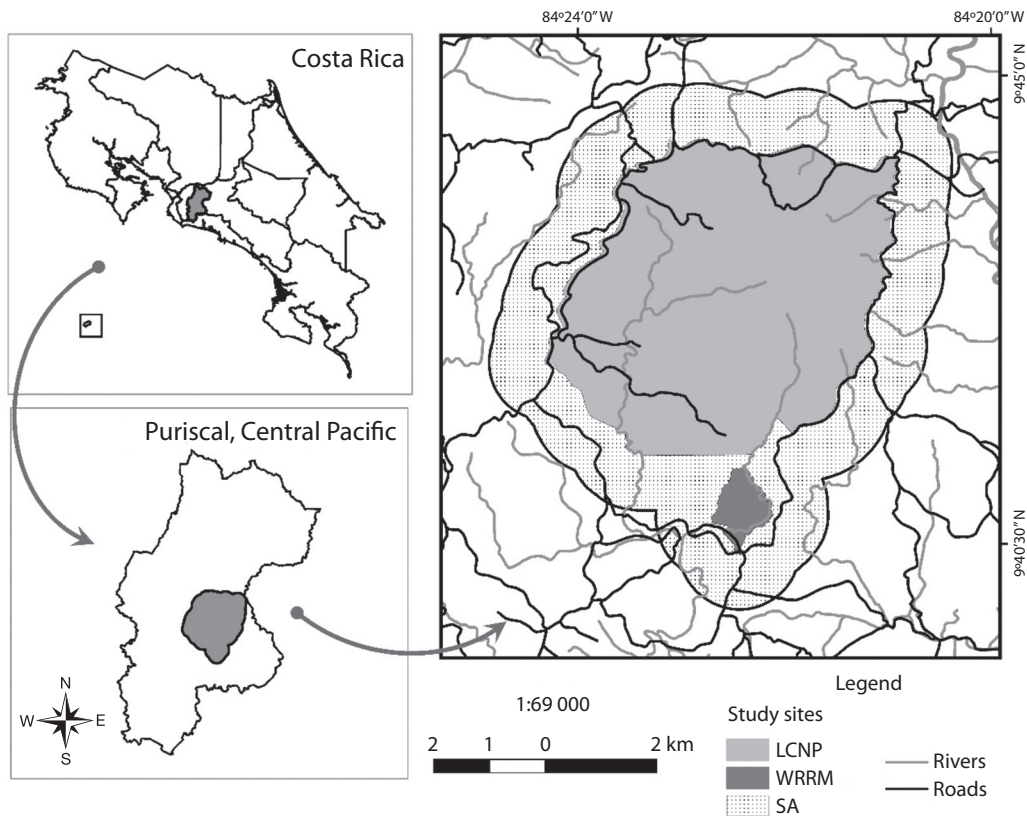


Fig. 1. Three landscapes studied in the Central Pacific of Costa Rica. La Cangreja National Park (LCNP), Rancho Mastatal Wildlife Refuge (WRRM), and surrounding area (SA).

are present in the landscape: altered *versus* natural coverage. From this, altered coverage was classified in four elements: crops as land dedicated to agriculture, pastures as areas dominated by herbaceous cover, open lands as areas without vegetation, and human settlements as any type of human infrastructure. Likewise, natural coverage was classified in three elements: forest cover as natural and native coverage that has not been tampered with or modified by humans, thickets as natural cover that had slight degree of human intervention, and water bodies as water mains like rivers, creeks and streams.

Different landscape metrics were extracted in all years of the study in order to characterize what the composition and configuration of the elements were at each delimited area and land use level. These metrics were obtained from

each thematic layer and from each delimited area according to Botequilha & Ahern (2002). In essence, for each year the number, average size, and average shape index of each fragment at the delimited area and land use level were determined. Area proportion of forest cover, thickets and pastures were determined within each area for the three years. These metrics were extracted by Patch Analyst extension from ArcGis 10 software (Rempel, Kaukinen, & Carr, 2012).

The methods of Peña-Cortés et al. (2006) were used to analyze the speed of change of each land use by the annual rate of land use change (RC) within the three areas: $RC = (A_f - A_i) / \Delta t$, where A is the calculated area (ha) for land use at the period f and i , and Δt is the time between in each period evaluated (1997-2005, 2005-2010 and 1997-2010).

Shannon Diversity (SDI) and Equity (SEI) indexes (Rempel et al., 2012) were applied to a set of land use subsamples in each of the delimited areas to determine the diversity and equity of elements as well as any pattern of variability. These were obtained through the creation of 25 ha hexagon plots on each delimited area to extract and apply later the land uses within each hexagon and diversity-equity index, respectively. Hexagon subsamples were created by Patch Analyst (Rempel et al., 2012). Because not all subsamples represented 25 ha, only subsamples greater than 25 % (6.25 ha) of area were used. From these, two Repeated Measures analyses on diversity and equity indices were performed to determine if differences existed between delimited areas, study periods, and their interaction.

Finally, to identify which regions in the whole landscape presented changes in land use, cross table analyses (Crosstab) were performed according to Bocco, Mendoza and Masera (2001) and Franco, Regil, González and Nava (2006). For this, the thematic layers of each year were transformed to raster format with a cell size of 25 m, and then cross-tabulations from raster layers between 1997-2005 and 2005-2010 were made. From this, a map of land use changes, a transition matrix with probabilities of land use changes, and their dependence of land use transition were obtained. This analysis was performed by IDRISI selva software (Eastman, 2012).

RESULTS

For the evaluated period, land use showed that the delimited areas were immersed in a natural matrix composed mainly by forest and thicket cover (Fig. 2). In general, the whole landscape in 1997 was comprised by 74 % natural cover and 26 % altered cover. However, in 2005 and 2010 the natural and altered cover increased and decrease by 1 %, respectively. Approximately 53 ha changed from altered to natural cover from 1997 to 2005 in the whole-landscape. These natural cover percentages corresponded mainly to forest cover (1997 =

52 %; 2005 = 54 %; 2010 = 57 %), while altered cover percentages corresponded mainly to pastures (1997 = 24 %; 2005 = 25 %; 2010 = 24 %). Forest cover areas were the most prevalent type of cover per period per site (Fig. 3), tending to increase slightly in the protected areas during 2005 and 2010, but not in SA. In contrast, thicket cover decreased between 2005 and 2010 at each site. Increases in pasture areas occurred only in SA between 2005 and 2010 when compared with 1997 (Fig. 3). Other land uses (human settlements, water bodies, crops and open lands) showed lower cover close to 2%, which were scattered mainly in SA (Fig. 2).

In relation to landscape metrics at the study area level, LCPN and SA showed an increase and a decrease in the number of fragments in 2005 and 2010, respectively, while their average size decreased and increased in the same way (Table 1). Pattern of changes in the shape index for the study sites were different, LCPN showed an increase towards 2010, RMWR showed a decrease, while SA presented a decrease and increase in 2005 and 2010, respectively (Table 1). Shannon Diversity and Equity indexes were different for the evaluated regions; SA had the highest values of diversity and equity independent of the period, due to the complexity of its elements; it was followed by LCPN and RMWR (Table 1). Both indices decreased towards 2010 for all three areas. Repeated measures analysis showed that there is a temporal effect over Shannon index ($F_{2,244} = 2.15$, $P = 0.005$) but not on the Equity of elements ($F_{2,244} = 0.02$, $P = 0.13$); besides, this analysis showed that an effect exists with the site-temporal interaction on Shannon ($F_{4,488} = 3.5$, $P < 0.001$) and Equity index ($F_{4,488} = 2.29$, $P = 0.004$).

Landscape metrics at land use level reflected that some land uses in LCPN and RMWR such as thickets and forest covers, decreased and increased in 2010, respectively, while in SA, human settlements and pastures coverage increased and forest cover decreased (Table 2). In general, the number of fragments decreased in all class uses except thickets and human

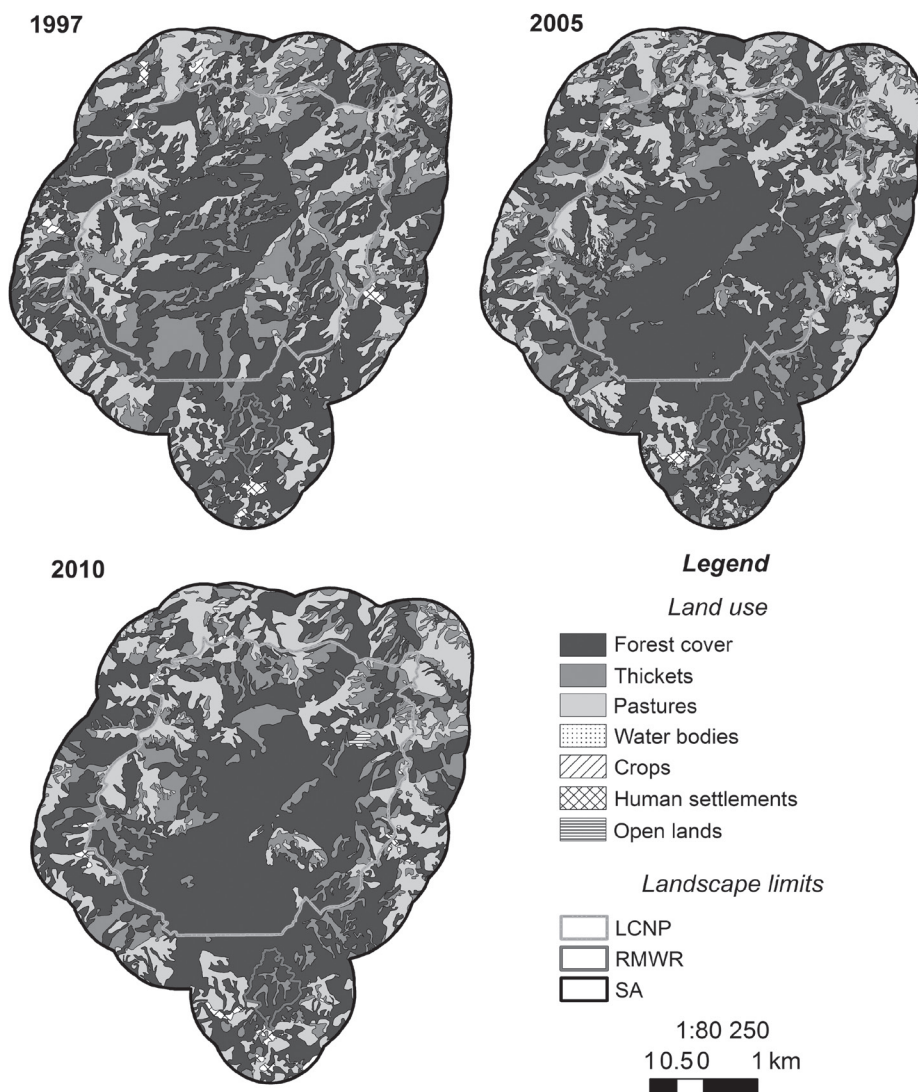


Fig. 2. Land use of three landscapes studied in the Central Pacific of Costa Rica in 1997, 2005 and 2010. La Cangreja National Park (LCNP), Rancho Mastatal Wildlife Refuge (RMWR), and surrounding area (SA).

settlements in 2005, but this trend was not observed in 2010 (Table 2). Moreover, average fragment size and shape index of each land use were different (Table 2), and thickets and forest cover within each area decreased and increased their size and shape from 1997 to 2010, respectively.

The speed of land use change annual rates for each study site was different depending on the year and the site area (Table 3). For

LCNP and RMWR, it was found that thickets and pastures presented negative rates of change between 1997 and 2010, while forests cover showed positive rates (Table 3). In SA, only human settlements and pastures presented positive annual rates of change between 1997 and 2010.

Crosstab analysis between land coverage of 1997-2005 and 2005-2010 showed that the whole landscape presented a dependence to

TABLE 1
Structural attributes of land use elements in the Central Pacific of Costa Rica in 1997, 2005 and 2010

Attributes	LCNP			Landscape RMWR			SA		
	1997	2005	2010	1997	2005	2010	1997	2005	2010
NumP	172	219	143	11	6	10	313	369	305
MPS (DS)	14.59 (83.52)	11.45 (104.63)	17.55 (132.93)	7.20 (11.20)	13.19 (22.20)	7.92 (18.87)	8.70 (31.99)	7.36 (56.54)	8.91 (65.09)
MSI	2.29	2.14	2.06	2.06	2.52	2.90	2.18	2.00	2.41
SDI	0.78 (0.25)	0.65 (0.37)	0.61 (0.38)	0.70 (0.14)	0.51 (0.13)	0.49 (0.15)	0.84 (0.26)	0.84 (0.28)	0.78 (0.27)
SEI	0.74 (0.21)	0.61 (0.32)	0.59 (0.34)	0.67 (0.16)	0.63 (0.19)	0.59 (0.19)	0.75 (0.19)	0.74 (0.21)	0.73 (0.22)

Abbreviations: La Cangreja National Park (LCNP), Rancho Mastatal Wildlife Refuge (RMWR), and surrounding area (SA). Number of patches (NumP), average size of patches (MPS), average shape index (MSI), Shannon diversity index (SDI), shannon equity index (SEI). Values in parentheses represent the standard deviation.

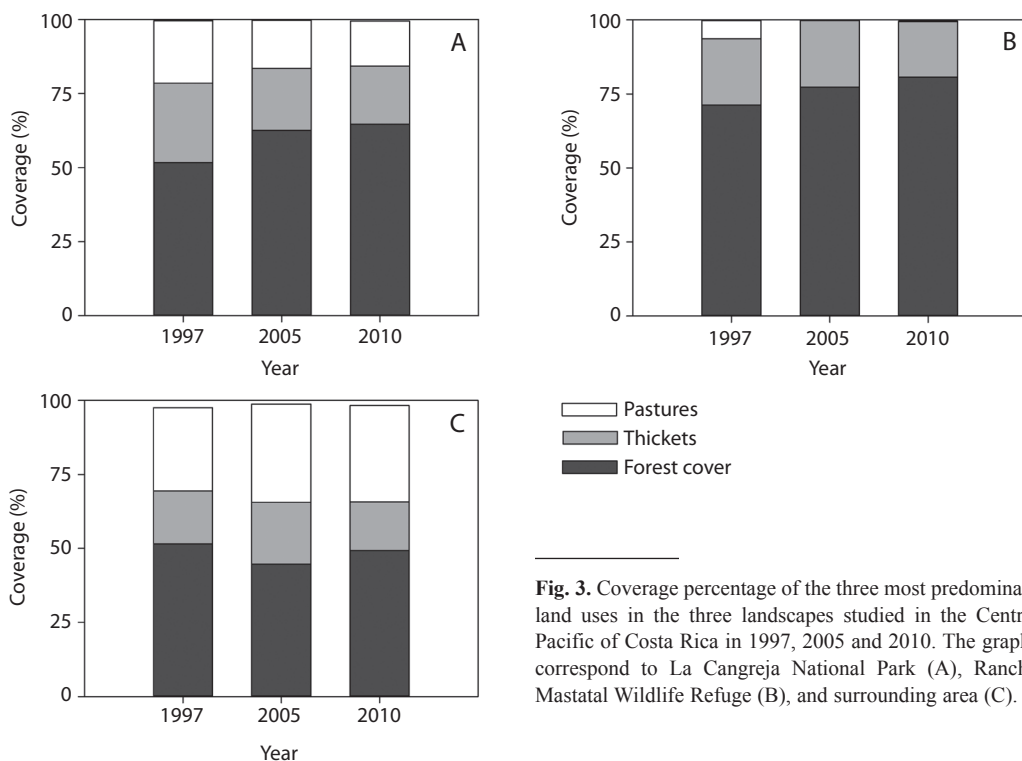


Fig. 3. Coverage percentage of the three most predominant land uses in the three landscapes studied in the Central Pacific of Costa Rica in 1997, 2005 and 2010. The graphs correspond to La Cangreja National Park (A), Rancho Mastatal Wildlife Refuge (B), and surrounding area (C).

change the land use and maintaining their use (1997-2005: $\chi^2_{36} = 10104372.8$, $P < 0.001$; 2005-2010: $\chi^2_{36} = 22330379.2$, $P < 0.001$). In general, the 1997-2005 and 2005-2010 transition matrix showed that forest cover, water bodies and pastures coverage presented greater probability of maintaining its previous state (0.49, 0.49 and 0.30, respectively), while forest cover, thickets and pastures coverage presented greater probability to change to other uses

(Table 4). The map of land use change (Fig. 4) showed that Northeastern area of SA presented a trend to change to pastures cover, while areas within LCNP and RMWR tend to change to forest cover.

DISCUSSION

In general, our results demonstrated that the natural matrix on the landscapes studied

TABLE 2
Spatial pattern characteristics at class level of the three landscapes study in the Central Pacific of Costa Rica in 1997, 2005 and 2010

Land use	Site	Cover (ha)			NumP			MPS (±DS) (ha)			MSI		
		1997	2005	2010	1997	2005	2010	1997	2005	2010	1997	2005	2010
Human settlements	LCNP	2.34	5.93	4.43	2	9	9	1.17±1.15	0.66±0.94	0.49±0.58	1.52	1.34	1.51
	RMWR	0.08	0.05	0.37	1	1	3	0.08±0.00	0.05±0.00	0.12±0.12	1.47	1.46	3.10
	SA	31.02	25.80	35.79	5	38	33	6.20±3.74	0.68±1.44	1.08±1.63	1.90	1.39	1.98
Forest cover	LCNP	1 297.58	1 569.85	1 621.70	33	12	12	39.32±186.01	130.82±428.15	135.14±440.74	2.29	2.73	2.38
	RMWR	56.43	61.19	63.88	4	1	1	14.11±15.65	61.19±0.00	63.88±0.00	2.06	3.94	3.62
	SA	1 402.45	1 212.38	1 338.92	76	32	35	18.45±61.48	37.89±186.24	38.25±187.31	2.23	2.30	3.63
Thickets	LCNP	673.28	523.50	492.72	65	126	69	10.36±19.05	4.15±7.87	7.14±10.13	2.46	2.15	2.15
	RMWR	17.71	17.91	14.85	2	3	4	8.85±2.23	5.97±7.07	3.71±3.62±	3.18	2.34	3.30
	SA	487.93	569.96	449.20	97	175	131	5.03±6.81	3.26±6.75	3.43±6.10	2.28	2.10	2.34
Water bodies	LCNP	0	0	0	0	0	0	0	0	0	0	0	0
	RMWR	0	0	0	0	0	0	0	0	0	0	0	0
	SA	1.96	1.92	1.34	1	1	1	1.96±0.00	1.92±0.00	1.34±0.00	2.00	1.91	1.97
Crops	LCNP	0.36	0	0	1	0	0	0.36±0.00	0	0	1.46	0	0
	RMWR	0	0	0	1	0	0	0	0	0	0	0	0
	SA	18.38	0.80	3.42	5	3	1	3.68±1.95	0.27±0.19	3.42±0.00	1.56	1.23	1.44
Pastures	LCNP	528.87	408.60	383.07	68	71	52	7.78±11.54	5.75±11.10	7.37±11.85	2.18	2.14	1.98
	RMWR	4.96	0.02	0.07	4	1	2	1.24±0.36	0.02±0.00	0.03±0.02	1.65	2.71	1.46
	SA	764.57	899.23	884.21	128	102	96	5.97±12.07	8.82±15.95	9.21±15.37	2.12	2.08	2.30
Open lands	LCNP	7.44	0.30	7.97	3	1	1	2.48±1.77	0.30±0.00	7.97±0.00	1.98	1.25	1.57
	RMWR	0	0	0	0	0	0	0	0	0	0	0	0
	SA	16.02	4.92	5.53	1	18	8	16.02±0.00	0.27±0.37	0.69±1.07±	1.47	1.50	1.61

Abbreviations: La Cangreja National Park (LCNP), Rancho Mastatal Wildlife Refuge (RMWR), and surrounding area (SA). Number of patches (NumP), mean patch size (MPS) and average shape index (MSI). Error values represent the standard deviation.

TABLE 3
Annual rate of land use change (ha/year) of three landscape study areas
in the Central Pacific of Costa Rica in 1997, 2005 and 2010

Land use	Landscape								
	LCNP			RMWR			SA		
	1997-2005	2005-2010	1997-2010	1997-2005	2005-2010	1997-2010	1997-2005	2005-2010	1997-2010
Human settlements	0.45	-0.30	0.16	0.00	0.06	0.02	-0.65	2.00	0.37
Forest cover	34.03	10.37	24.93	0.60	0.54	0.57	-23.76	25.31	-4.89
Thickets	-18.72	-6.16	-13.89	0.02	-0.61	-0.22	10.25	-24.15	-2.98
Water bodies	---	---	---	---	---	---	0.00	-0.12	-0.05
Crops	-0.05	0.00	-0.03	---	---	---	-2.20	0.52	-1.15
Pastures	-15.03	-5.11	-11.22	-0.62	0.01	-0.38	16.83	-3.01	9.20
Open lands	-0.89	1.53	0.04	0.00	0.00	0.00	-1.39	0.12	-0.81

Abbreviations: La Cangreja National Park (LCNP), Rancho Mastatal Wildlife Refuge (RMWR), and surrounding area (SA).

TABLE 4
Probability matrix of land use transition and stability for the whole landscape studied
in the Central Pacific of Costa Rica between 1997-2005 and 2005-2010

		Probability matrix 1997-2005						
		2005						
		Human settlements	Forest cover	Thickets	Water bodies	Crops	Pastures	Open lands
1997	Human settlements	0.04	0	0	0	0	0	0
	Forest cover	0	0.49	0.11	0	0	0.02	0
	Thickets	0	0.07	0.07	0	0	0.08	0
	Water bodies	0	0	0	0.49	0	0	0
	Crops	0	0	0	0	0	0	0
	Pastures	0.01	0.03	0.04	0	0	0.30	0
	Open lands	0	0	0	0	0	0	0
		Probability matrix 2005-2010						
		2010						
		Human settlements	Forest cover	Thickets	Water bodies	Crops	Pastures	Open lands
2005	Human settlements	0.25	0	0	0	0	0	0
	Forest cover	0	0.76	0.05	0	0	0	0
	Thickets	0	0	0.28	0	0	0.03	0
	Water bodies	0	0	0	0.65	0	0	0
	Crops	0	0	0	0	0	0	0
	Pastures	0	0	0.01	0	0	0.63	0
	Open lands	0	0	0	0	0	0	0.07

White cells represent the probability that land use will transition to another use, and gray cells represent the probability that land use will remain stable.

were mainly composed of forest and thickets cover. This conclusion was based on three facts: **1)** these were the most dominant in the landscape, **2)** they were the most connected, and **3)** they were the most prevalent coverage in temporal terms. Although there is a higher matrix of natural cover on the whole landscape, the existence or absence of an anthropogenic effect reflects the dynamics and structure of landscape elements in each study site, as

generally described in other landscape analyses by Peña-Cortés et al. (2006). In terms of land use through time, decrease in the number of fragments and shape index, and increase or decrease in size of patches can be an indication of landscape evolution in the area. In relation to the latter, increases in forest cover and pastures inside and outside of protected areas, can be seen as indicative of elements that dominate the spatial and temporal landscape dynamics of the



Fig. 4. Land use change of the three study areas in the Central Pacific of Costa Rica between 1997-2005 and 2005-2010. La Cangreja National Park (LCNP), Rancho Mastatal Wildlife Refuge (RMWR), and surrounding area (SA).

region, which may be the result of the protected areas conservation efforts and livestock activity in the zone.

The pattern in all years of great number of fragments, diversity and equity of land use founded in SA compared with LCNP and RMWR demonstrates the high landscape complexity in the edges of protected areas. Turner (1989, 2005) showed that regions with greater number of elements, fragments and less natural cover usually describe areas with high landscape complexity, which are generally susceptible to disturbance and habitat loss. Evidence of this phenomenon in the Central Pacific of Costa Rica are observed in Manuel Antonio National Park, Carara National Park and other areas near to the coast, where the urban growth and different crops around the protected areas generate a decline of wildlife populations (Broadbent et al., 2012; Sáenz & Sáenz, 2007; Sanchez-Azofeifa et al., 2001).

Changes in land use within LCNP and RMWR show the recovery of forested areas likely due to conservation efforts in the zone. It has been shown that remaining structure of forest cover in the landscape is important for the recovery of areas where the tropical forest did not exist (Helmer, 2000). In this sense, it is likely that the remaining coverage prior to the creation of these protected areas was

responsible for the recovery of thickets and forest cover. Likewise, the great coverage, low number of fragments and high stability of forest cover within the two protected areas, show the potential for conservation of natural habitat that these protected zones have. In general, forest cover can play the role of core areas for endemic wildlife unique species to the area, such as *Plinia puriscalensis* and species with slow growth such as *Aspidosperma myrasticifolium*, that require forest cover for development and growth (Guzmán & Cordero, 2013). Furthermore, forest cover recovery within LCNP and RMWR must be considered as positive aspects for tourism attraction (Alvarado, 2008; Díaz, van Koppen, Breitling, & de Camino, 2005).

Considering the observed changes in SA, it is likely that problems of the surrounding communities, such as deforestation and burning of land for the generation of private farms (Jiménez, 2011), are responsible for pasture cover increases. In general, Northeastern and Eastern areas outside and the Northern areas inside of the LCNP were the areas where the large land use changes to altered coverage occurred. This may indicate the large anthropogenic pressures that experience these zones. In terms of habitat connectivity, it is possible that some wildlife species of LCNP are suffering geographical isolation. This can be evidenced because the

Northern, Western and in the lower portion the Eastern regions did not have forest cover continuity and showed high altered coverage, which probably limited the species distribution that require this unaltered habitat. Although some studies promote the payment for environmental services to increase connectivity between protected areas (Sánchez-Azofeifa et al., 2003; Sánchez-Azofeifa, Pfaff, Robalino, & Boomhower, 2007), it is clear that these measures cannot achieve their goal if the local governments promote or do not supervise the anthropogenic changes around the protected areas. For this reason, local governments must contemplate future policies of connectivity in the zone with appropriate land management strategies that considers social-environmental aspects to promote the conservation and management in protected areas.

The case of LCNP and RMWR reflects that these recently designated protected areas have showed rapid regeneration over 13 years. However, the current anthropogenic pressure in SA and lack of effectiveness of local governments to monitor and abate the land use changes have caused a shift in land use to a condition of altered cover, which can potentially hinder the management, conservation and restorations of species in the protected areas. We emphasize the fact that these protected areas in the Central Pacific of Costa Rica are entities of conservation and restoration of the forest cover.

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RESUMEN

¿Es la cobertura forestal conservada y restaurada por las zonas protegidas?: El caso de dos áreas silvestres protegidas en el Pacífico Central de Costa Rica. Cambios en el uso del suelo son principalmente consecuencia de las acciones antropogénicas. La actual transformación agrícola y urbana en Costa Rica ha generado preguntas acerca de la efectividad de la conservación y restauración dentro de las áreas protegidas. En este documento nosotros analizamos los patrones de cambio del uso del suelo entre tres periodos: 1997, 2005 y 2010 en términos de magnitud, dirección y velocidad a través de mapas categóricos generados por la foto-interpretación para dos áreas silvestres protegidas y sus áreas aledañas: Parque Nacional La Cangreja (LCNP), el Refugio de Vida Silvestre Rancho Mastatal (RMWR) y sus áreas aledañas (SA), esta última compuesta por área de un kilómetro de radio fuera de los límites de las áreas protegidas. La matriz que describe el paisaje dentro de las áreas protegidas es la cobertura natural, compuesta principalmente por la cobertura forestal y tacotales. Encontramos que la cobertura natural más abundante para ambas áreas protegidas fue cubierta forestal en todos los años estudiados. La estabilidad y las grandes áreas de la cubierta forestal en LCNP y RMWR para 2005 y 2010 reflejan que las políticas, las acciones de manejo y vigilancia tienen un impacto positivo en la conservación y restauración de los hábitats naturales en esta zona del Pacífico Central Costarricense. Sin embargo, la alta complejidad del paisaje de SA en 1997, 2005 y 2010 son una prueba de presión antropogénica sobre estas áreas protegidas y sugieren una ineficacia de los gobiernos locales para monitorear y disminuir los cambios de uso del suelo que podrían obstaculizar la gestión, conservación y restauración de especies dentro de las áreas protegidas.

Palabras clave: Parques Nacionales de Costa Rica, fragmentación de hábitat, ecología del paisaje, cambio del uso del suelo, Parque Nacional La Cangreja.

REFERENCES

- Acosta-Vargas, L. G. (1998). *Análisis de la composición florística y estructura para la vegetación del piso basal de la zona protectora La Cangreja, Mastatal de Puriscal*. (Specialty Practice Report). Instituto Tecnológico de Costa Rica, Cartago, Costa Rica.
- Aguayo, M., Pauchard, A., Azócar, G., & Parra, O. (2009). Cambio del uso del suelo en el centro sur de Chile a fines del siglo XX: Entendiendo la dinámica espacial y temporal del paisaje. *Revista Chilena de Historia Natural*, 82(3), 361-374. doi:10.4067/S0716-078X2009000300004.
- Alvarado, L. A. (2008). *Propuesta de sostenibilidad turística para ser desarrollada en el Parque Nacional La Cangreja, (PNLC), ubicado en Puriscal, San Jose*.



- Costa Rica. (Masters thesis). Universidad para la Cooperación Internacional, San José, Costa Rica.
- Bender, O., Boehmer, H. J., Jens, D., & Schumacher, K. P. (2005). Using GIS to analyse long-term cultural landscape change in Southern Germany. *Landscape and Urban Planning*, 70(1), 111-125. doi:10.1016/j.landurbplan.2003.10.008.
- Bocco, G., Mendoza, M., & Masera, O. R. (2001). La dinámica del cambio del uso del suelo en Michoacán: Una propuesta metodológica para el estudio de los procesos de deforestación. *Investigaciones Geográficas*, 44, 18-36.
- Bonilla, A. (1983). Proceso histórico de los recursos naturales en el Cantón de Puriscal. In J. Heuvelodp & L. Espinoza (Eds.). *El componente arbóreo en Acosta y Puriscal, Costa Rica* (pp. 11-17). San José, Costa Rica: CATIE.
- Botequilha, A. & Ahern, J. (2002). Applying landscape ecological concepts and metrics in sustainable landscape planning. *Landscape and Urban Planning*, 59(2), 65-93. doi:10.1016/S0169-2046(02)00005-1.
- Broadbent, E. N., Zambrano, A. M. A., Dirzo, R., Durham, W. H., Driscoll, L., Gallagher, P., Salters, R., Schultz, J., Colmenares, A., & Randolph, S. G. (2012). The effect of land use change and ecotourism on biodiversity: A case study of Manuel Antonio, Costa Rica, from 1985 to 2008. *Landscape Ecology*, 27(5), 731-744. doi:10.1007/s10980-012-9722-7.
- Bustamante, J. 2006. *Corredor Biológico Paso de las Lapas* (Technical report). Sistema Nacional de Áreas de Conservación, Costa Rica.
- Costanzo, A. J. (2006). *A quantitative survey of riparian forest structure along the Quebrada Grande in La Cangreja National Park, Costa Rica*. (Masters thesis). University of Washington, USA.
- Díaz, E., van Koppen, K., Breitling, J., & de Camino, R. (2005). Ecoturismo y desarrollo rural en el Parque Nacional La Cangreja, Costa Rica. *Recursos Naturales y Ambiente*, 45, 120-126.
- Eastman, J. R. (2012). *IDRISI selva: manual*. USA: Clark Lab, Clark University, Worcester MA.
- Fahrig, L. (2003). Effects on habitat fragmentation on biodiversity. *Annual Review of Ecology and Systematics*, 34, 487-515.
- Franco, S., Regil, H., González, C., & Nava, G. (2006). Cambio de uso del suelo y vegetación en el Parque Nacional Nevado de Toluca, México, en el periodo 1972-2000. *Investigaciones Geográficas*, 61, 38-57.
- Guzmán, J., & Cordero, R. (2013). Growth and photosynthetic performance of five tree seedlings species in response to natural light regimes from the Central Pacific of Costa Rica. *Revista de Biología Tropical*, 61(3), 1433-1444. doi:http://dx.doi.org/10.15517/rbt.v61i3.11970.
- Haila, Y. (2002). A conceptual genealogy of fragmentation research: from island biogeography to landscape ecology. *Ecological Applications*, 12(2), 321-334. doi:10.1890/1051-0761(2002)012[0321:ACGOFR]2.0.CO;2.
- Hanski, I. (1998). Metapopulation dynamics. *Nature*, 396(6706), 41-49. doi:10.1038/23876.
- Helmer, E. H. (2000). The landscape ecology of tropical secondary forest in motane Costa Rica. *Ecosystems*, 3(1), 98-114. doi:10.1007/s100210000013.
- Holdridge, L. R. (1967). *Life zone ecology*. San José, Costa Rica: Tropical Science Center.
- Jiménez, Q. (1999). *Importancia biológica de la zona protectora La Cangreja, Puriscal*. (Technical report). Heredia, Costa Rica.
- Jiménez, V. (2011). Problemática ambiental del Parque Nacional La Cangreja y de las comunidades aledañas. *Biocenosis*, 25(1-2), 5-19.
- McLennan, B., & Garvin, T. (2012). Intra-regional variation in land use and livelihood change during a forest transition in Costa Rica's dry North West. *Land Use Policy*, 29(1), 119-130. doi:10.1016/j.landusepol.2011.05.011.
- Myers, N. (1993). Tropical forests: The main deforestation fronts. *Environmental Conservation* 20(1): 9-16. doi:10.1017/S0376892900037176.
- Peña-Cortés, F., Rebolledo, G., Hermosilla, K., Hauensstein, E., Bertrán, C., Schlatter, R., & Tapia, J. (2006). Dinámica del paisaje para el período 1980-2004 en la cuenca costera del Lago Budi, Chile. Consideraciones para la conservación de sus humedales. *Ecología Austral*, 16(2), 183-196.
- Poder Ejecutivo. (2004). *Decreto ejecutivo N° 31631-MINAE. 10-02-2004*. San José, Costa Rica: La Gaceta.
- Poder Ejecutivo. (2005). *Decreto ejecutivo N° 32752-MINAE. 20-06-2005*. San José, Costa Rica: La Gaceta.
- Rempel, R. S., Kaukinen, D., & Carr, A. P. (2012). *Patch Analyst and Patch Grid*. Ontario Ministry of Natural Resources, Centre for Northern Forest Ecosystem Research, Thunder Bay, Ontario.
- Rosero-Bixby, L., & Palloni, A. (1998). Population and deforestation in Costa Rica. *Population and Environment*, 20(2), 149-185. doi:10.1023/a:1023319327838.
- Sáenz, J. & Sáenz P. (2007). Influencia de las variables de hábitat y paisaje sobre la presencia-ausencia del mono Titi y el mono Carablanca en un área fragmentada del Pacífico Central de Costa Rica. In C. A. Harvey, & J. C. Sáenz (Eds.). *Evaluación y Conservación de la Biodiversidad en Paisajes Fragmentados de Mesoamérica* (pp. 511-545). Santo Domingo de Heredia, Costa Rica: Editorial INBio.

- Sánchez-Azofeifa, A., Daily, G. C., Pfaff, A. S. P., & Busch, C. (2003). Integrity and isolation of Costa Rica's national parks and biological reserves: examining the dynamics of land-cover change. *Biological Conservation*, *109*(1), 123-135. doi:10.1016/S0006-3207(02)00145-3.
- Sanchez-Azofeifa, G. A., Harriss, R. C., & Skole, D. L. (2001). Deforestation in Costa Rica: A Quantitative Analysis Using Remote Sensing Imagery1. *Biotropica*, *33*(3), 378-384. doi:10.1111/j.1744-7429.2001.tb00192.x.
- Sánchez-Azofeifa, G. A., Pfaff, A., Robalino, J. A., & Boomhower, J. P. (2007). Costa Rica's payment for environmental services program: Intention, implementation, and impact. *Conservation Biology*, *21*(5), 1165-1173. doi:10.1111/j.1523-1739.2007.00751.x.
- Sanchez-Azofeifa, G. A., Rivard, B., Calvo, J., & Moorthy, I. (2002). Dynamics of Tropical Deforestation Around National Parks: Remote Sensing of Forest Change on the Osa Peninsula of Costa Rica. *Mountain Research and Development* *22*(4): 352-358. doi:10.1659/0276-4741(2002)022[0352:DOTDAN]2.0.CO;2.
- SINAC. (2010). *Sistema nacional de áreas de conservación: Plan estratégico quinquenio 2010-2015* (Technical Report). San José, Costa Rica: Ministerio de Ambiente, Energía y Telecomunicaciones.
- Turner, M. (1989). Landscape ecology: The effect of pattern on process. *Annual Review of Ecology and Systematic*, *20*, 171-191.
- Turner, M. (2005). Landscape ecology: What is the state of the science? *Annual Review of Ecology and Systematic*, *36*, 319-344.
- Wiens, J. A., Stenseth, N. C., Van Horne, B., & Anker, R. (1993). Ecological mechanisms and landscape ecology. *Oikos*, *66*, 369-380.

